

APPENDIX 7. PROCESSES THREATENING AUSTRALIA'S ENDANGERED AND VULNERABLE TERRESTRIAL REPTILES.

Most of these threats are speculative; few have been confirmed by research. See individual recovery outlines for further explanation of the processes which are believed to be responsible for declines.

SPECIES	THREATENING PROCESS															Other threats			
	Habitat clearance	Over-grazing by stock	Cropping	Pre- dation	Urban develop- ment	Pasture improve- ment	Fire regime	Soil degrad- ation	Visitor disturbance	Soil &/or water pollution	Mining	Native forest logging	Climatic vari- ation	Rabbit grazing	Habitat fragment- ation		Weed invasion	Habitat drainage	Rock removal
<i>Eiaseya</i> sp. nov.	*							*		*							? *		
<i>Emydura signata</i>								*		*		*							Fishing ?water supply scheme
<i>Pseudemydura umbrina</i>	*		*	*	*		*				*		*				*		
<i>Rheodytes leukops</i>	*		*							*	*								
Gen. nov. sp. nov. (Chelidae)	*	*	*	*					*	*	*								Irrigation Captive trade
<i>Christinus guentheri</i>				*				*						*					
<i>Lepidodactylus listeri</i>	*			*							*					*			Displaced by introduced sp.
<i>Nephurus deileani</i>		*						*						*					
<i>Underwoodisaurus sphyrurus</i>	*	*										*							Dam building
<i>Aprasia aurita</i>	*		*							*									
<i>Aprasia pseudopuichella</i>	*	*	*		*					*									
<i>Aprasia rostrata rostrata</i>				*					*										
<i>Delma impar</i>	*	*	*		*			*		*					*	*	*	*	*

SPECIES	THREATENING PROCESS																			
	Habitat clearance	Over-grazing by stock	Cropping	Pre-dation	Urban development	Pasture improvement	Fire regime	Soil degradation	Visitor disturbance	Soil &/or water pollution	Mining	Native forest logging	Climatic variation	Rabbit grazing	Habitat fragmentation	Weed invasion	Habitat drainage	Rock removal	Other threats	
<i>Delma labialis</i>	*	*		*					*											
<i>Delma torquata</i>	*	*		*	*															
<i>Ophidiocephalus taeniatus</i>		*					*						*	*					Loss of litter	
<i>Paradelma orientalis</i>	*	*	*		*							*								
<i>Ctenophorus yinnietharra</i>		*																		
<i>Tympanocryptis lineata pinguicollis</i>	*	*	*		*									*	*	*		*	Irrigation	
<i>Anomalopus mackayi</i>	*	*	*				*	*		*									Loss of litter Irrigation	
<i>Coeranoscincus reticulatus</i>	*	*	*			*						*			*					
<i>Ctenotus lanceolini</i>									*							*			Disturbance by gull colony	
<i>Ctenotus zastictus</i>		*							?	*										
<i>Egernia kintorei</i>				*			*			*									Hunted for food	
<i>Egernia pulchra longicauda</i>				*			*		*										Competition from introd. sp.	
<i>Egernia stokesii aethiops</i>				*									*						Competition from native sp.	
<i>Egernia stokesii badia</i>	*	*	*																	

THREATENING PROCESS																			
SPECIES	Habitat clearance	Over-grazing by stock	Cropping	Pre-dation	Urban development	Pasture improvement	Fire regime	Soil degradation	Visitor disturbance	Soil &/or water pollution	Mining	Native forest logging	Climatic variation	Rabbit grazing	Habitat fragmentation	Weed invasion	Habitat drainage	Rock removal	Other threats
<i>Egernia stokesii stokesii</i>				*									*						
<i>Eulamprus leuraensis</i>				?*	*				*	*		*							Plantations
<i>Eulamprus tympanum ssp. nov.</i>	*																	*	
<i>Lerista allanae</i>		*	*			*													
<i>Lerista vittata</i>	*					*													
<i>Niveoscincus palfreymani</i>				*									*						Loss of food source
<i>Pseudemoia lichenigera</i>				*				*						*					
<i>Tiliqua adelaidensis</i>			*		*														
<i>Ramphotyphlops exocoeti</i>	*							*			*								
<i>Aspidites ramsayi</i>	*		*	?*															
<i>Morelia spilota imbricata</i>	*	*	*			*													
<i>Austrelaps labialis</i>	*		*	*	*														
<i>Denisonia maculata</i>	*	*	*		*	*													?poisoned by introd. sp.
<i>Echiopsis atriceps</i>	*		*								*								
<i>Echiopsis curta</i>	*	*	*				?*												

THREATENING PROCESS

SPECIES	Habitat clearance	Over-grazing by stock	Cropping	Pre-emption	Urban development	Pasture improvement	Fire regime	Soil degradation	Visitor disturbance	Soil &/or water pollution	Mining	Native forest logging	Climatic variation	Rabbit grazing	Habitat fragmentation	Weed invasion	Habitat drainage	Rock removal	Other threats
<i>Elapognathus minor</i>	*				*							*					*		Competition from native sp.
<i>Furina dunmalli</i>	*	*	*		*	*											? *		
<i>Hoplocephalus bungaroides</i>	*				*				*						*			*	Captive trade
<i>Notechis ater ater</i>	*	*					*	*		*									? Loss of food source
<i>Simoselaps calanotus</i>	*		*		*										*	*			
TOTAL	30	21	21	14	14	12	10	9	8	7	6	6	5	6	5	5	4	4	

APPENDIX 8.1 WORLDMAP - Analysis of Priority Areas for Conservation of Reptiles

Threatened species represent a high-priority subset of the overall biodiversity of any given area. The measurement of biodiversity, whether in geographical, geo-taxonomic or simply taxonomic terms, continues to present biologists with a major problem.

In practice, biodiversity is usually measured by some simple index of taxonomic diversity such as species richness. But species richness is only one measure of diversity, and one which is often inappropriate to conservation action because it weights all species equally. Consequently phylogenetically-distinctive organisms (such as species representing monotypic families or orders) must be separately identified and given special emphasis in conservation where appropriate.

The supplemental variables in the Millsap *et al.* (1990) ranking system described in Appendix 10 attempt, with some success, to assign higher values to phylogenetically significant taxa. However in recent years a number of attempts have been made to utilise a wide range of variables (including phylogenetic significance) in identifying areas or habitats of special conservation significance and in assigning conservation priorities (eg Pressey *et al.*, 1993). All such methodologies eventually depend for their effectiveness on the quality of their underlying database, and on the rigour of their underlying mathematical model.

These problems have been very thoroughly addressed by Vane-Wright *et al.* (1991) and Williams *et al.* (1991, 1993). These authors point out that phylogenies based on cladistic methods provide an effective measure of phylogenetic uniqueness, and Williams has developed special computer software ("WORLDMAP") to measure various components of biodiversity and, through a series of elegant algorithms, to use various criteria to identify geographic areas (on a variety of scales) of high biodiversity conservation value and hence of potential value as reserves.

While this software is still in the development stage, the Consultants are indebted to Dr Paul Williams and his colleagues at the British Museum (Natural History) for making copies of their software available for trialing in the Australian Reptile Action Plan.

In order to exploit the full capabilities of

WORLDMAP it could be necessary to have available a cladistic analysis of the fauna being treated. However few groups of Australian reptiles have been subject to rigorous cladistic analysis.

Fortunately WORLDMAP currently uses only the *number of nodes* between the species and the root of the cladogram, and not *internodal distances*, as the principal measure of cladistic divergence. Consequently an equal-weight cladogram based on the current classification of the Australian reptile fauna provides the necessary cladistic input for a WORLDMAP analysis, and for the construction of the necessary cladistic codes.

The full WORLDMAP methodology, including the mathematical and statistical bases for the various analyses, is explained in the papers referred to above and should be consulted by the reader. The following brief descriptions of the nine principal biodiversity measures are extracted and only slightly modified from the WORLDMAP manual. As pointed out by Vane-Wright *et al.* (1991) no one measure can meet all of the criteria which need to be considered in setting biodiversity reserve priorities.

The following nine analyses were applied only to the 11 Endangered and 36 Vulnerable terrestrial reptiles.

Species richness

One of the simplest measures of diversity is unweighted species richness, an estimate of the number of species in an area.

The Species Richness maps (Appendices 5.2-5.4) display the unweighted numbers of species per square plotted as the raw numbers of species in the clade. The species richness map for marine turtles is shown in Appendix 6.1.

Endemism

While endemism is not a direct biodiversity measure, it represents an important criterion for conservation biologists seeking to prioritise protection measures for particular taxa. Centres of endemism can be identified by the co-occurrence of large numbers of species with very restricted distributions.

The Endemism map (Appendix 8.2) shows for Endangered and Vulnerable species combined,

the sum of the rarity scores (the inverse of the number of squares with records) for the species plotted as a percentage of the total rarity scores for all of the species in the clade.

"Close-to-root" species

Taxonomic root weighting measures diversity as species richness with higher weights for the early-diverging, "relict" species.

The Root Weight map (Appendix 8.3) displays for Endangered and Vulnerable species combined, the numbers of species per square, each weighted by its individual root weight, plotted as a percentage of the total root score for all of the species in the clade.

Mean root weight

Weighted faunal diversity scores can be thought of as consisting of two components: (A) the number of species per square, and (B) the mean weight per species in that square.

The Mean Root Weight map (Appendix 8.4) plots for Endangered and Vulnerable species combined, the weight component B and the mean root weighting per species for the fauna of each square as a percentage of the sum of the root weights of all of the species in the clade.

Higher-taxon richness

Higher-taxon richness weighting measures diversity as species richness with higher weights for faunas with more of the species from the more early-diverging, higher taxa, irrespective of the number of surviving species in these higher taxa.

The Higher Taxon Weighting map (Appendix 8.5) plots for Endangered and Vulnerable species combined, the numbers of species per square, weighted by the higher-taxon richness weight, as a percentage of the total higher-taxon richness score for all of the species in the clade.

Note that because higher-taxon richness is measured in terms of pairwise divergences, there must be at least two species in a square before its score can be greater than zero.

Mean higher taxonomic weight

Weighted faunal diversity scores can be thought of as consisting of two components: (A) the number of species per square, and (B) the mean weight per species in that square.

The Mean Higher Taxonomic Weight map (Appendix 8.6) plots for Endangered and Vulnerable species combined, the weight component B, the mean higher-taxon richness weighting per species for the fauna of each square as a percentage of the higher-taxon richness score for all of the species in the clade.

Spanning-tree length

Spanning-tree length weighting measures diversity as species richness, with higher weights for faunas with a greater variety of species as shown by the proportion of the classification represented.

The Span (unrooted-subtree) Weight map (Appendix 8.7) plots for Endangered and Vulnerable species combined, the numbers of species per square, weighted by the spanning-tree length weight, as a percentage of the total spanning-tree length score for all of the species in the clade.

Note that because spanning-tree length is measured in terms of the number of branching points between species, there must be at least two species in a square before its score can be greater than zero.

Mean spanning-tree weight

Weighted faunal diversity scores can be thought of as consisting of two components: (A) the number of species per square, and (B) the mean weight per species in that square.

The Mean Spanning Tree Weight map (Appendix 8.8) plots for Endangered and Vulnerable species combined, the weight component B, the mean spanning-tree length weighting per species for the fauna of each square, as a percentage of the entire spanning-tree length score for all of the species in the clade.

Taxonomic dispersion

Taxonomic dispersion weighting measures diversity as species richness, with higher weights for faunas with species that represent the variety of subgroups in the classification more evenly.

The Dispersion Weight map (Appendix 8.9) plots for Endangered and Vulnerable species combined, the numbers of species per square, weighted by their dispersion weight, as a percentage of the total dispersion score for all of the species in the clade.

Note that because cladistic dispersion is measured

in terms of the evenness of pairwise divergences, there must be at least three species in a square before its dispersion score can be greater than zero.

Mean taxonomic dispersion weight

Pairwise weighted faunal diversity scores can be thought of as consisting of two components: (A) the number of species per square, and (B) the mean pairwise weight per species in that square.

The Mean Dispersion Weight map (Appendix 8.10) plots for Endangered and Vulnerable species combined, the weight component B, the mean dispersion weighting per species for the fauna of each square, on the grid as a percentage of the dispersion score for all of the species in the clade.

Priority areas

Table 3 below shows the 13 areas (at a resolution of 2° of latitude and longitude) which rank highest in the various biodiversity analyses. In grid numerical order they are:

- 110 Christmas Island, Indian Ocean
- 345 Emerald, QLD
- 346 Rockhampton, QLD
- 383 Maryborough, QLD
- 418 Dalby, QLD
- 435 Houtman Abrolhos, WA
- 472 Perth, WA
- 473 Northam, WA
- 490 Armidale, NSW
- 520 Peterborough, SA
- 556 Adelaide, SA

- 593 Mt Gambier, SA & VIC
- 594 Ballarat, VIC

These areas therefore warrant special attention when designating new reserves or managing existing reserves to conserve Australia's Endangered and Vulnerable terrestrial reptiles.

References:

Pressey, R.L., Humphries, C.J., Margules, C.R., Vane-Wright, R.I. and Williams, P.H. 1993. Beyond opportunism: key principles for systematic reserve selection. *Trends in Ecology and Evolution* 8(4): 124-128.

Millsap, B.A., Gore, J.A., Runde, D.E. and Cerulean, S.I. 1990. Setting priorities for the conservation of fish and wildlife species in Florida. *Wildlife Monographs* 111: 1-57.

Vane-Wright, R.I., Humphries, C.J. and Williams, P.H. 1991. What to protect? - systematics and the agony of choice. *Biological Conservation* 55: 235-254.

Williams, P.H., Humphries, C.J. and Vane-Wright, R.I. 1991. Measuring biodiversity: taxonomic relatedness for conservation priorities. *Australian Systematic Botany* 4: 665-679.

Williams, P.H., Vane-Wright, R.I. and Humphries, C.J. 1993. Measuring biodiversity for choosing conservation areas. pp. 309-328 in J. LaSalle and I.D. Gauld (eds) *Hymenoptera and Biodiversity*. CAB International, Wallingford, UK.

TABLE 3 PRIORITY CONSERVATION AREAS (LISTED AS NUMBERED 2° GRID SQUARES) IDENTIFIED BY WORLDMAP'S NINE BIODIVERSITY MEASURES

BIODIVERSITY MEASURE	SEQUENCE OF PRIORITY FOR CONSERVATION AREAS			
	PRIORITY 1	PRIORITY 2	PRIORITY 3	PRIORITY 4
Species Richness	472	473	556,346,418	
Endemism	472	110	520	418
Root Weight	472	473	346	
Mean Root Weight	383	345,110		
Higher Taxon Weight	472	473	346	345
Mean Higher Taxon Weight	490	345	346	
Span Weight	472	473	594	556,346
Mean Span Weight	435, 593, 490	594		
Dispersion Weight	472	473	594	
Mean Dispersion Weight	594	435,593,490		

NUMBERS AND NAMES APPLIED TO THE 2° GRID SQUARES USED IN THE WORLDMAP ANALYSES

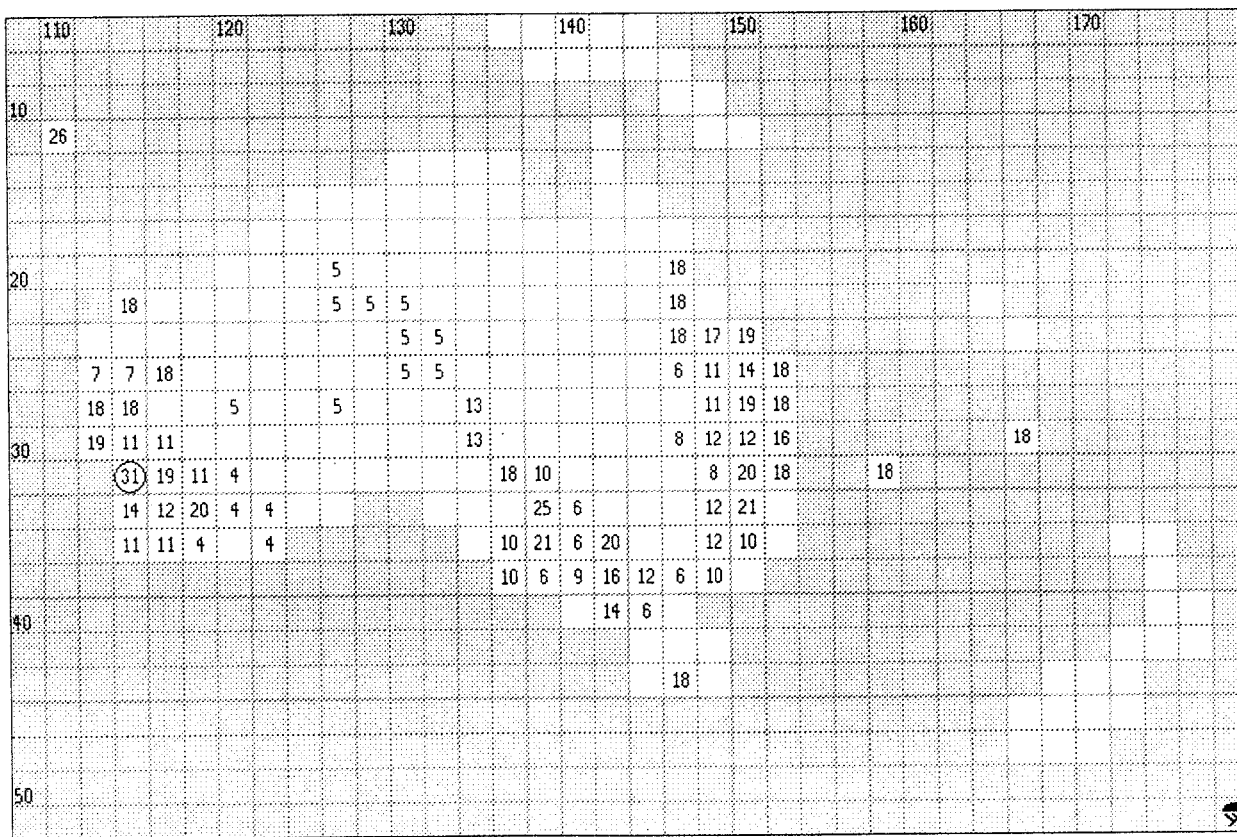


110	Christmas Island	193	Katherine	235	Cairns
120	Melville Island	194	Roper River	236	Innisfail
121	Cobourg Peninsula	195	Sir Edward Pellew	259	Lagrange
122	Cape Stewart	197	Edward River	260	Dampier Downs
123	Cape Wessel	198	Musgrave	261	Fitzroy Crossing
126	Cape York	199	Cooktown	262	Halls Creek
155	Cape Ford	224	Broome	263	Gardner Range
156	Darwin	225	King Leopold Ranges	264	Hooker Creek
157	Alligator Rivers	226	Durack Range	265	Renner Springs
158	Arnhem Land	227	Lake Argyle	266	Tennant Creek
159	Gove Peninsula	228	Victoria River Downs	267	Alexandria
161	Weipa	229	Newcastle Waters	268	Gunpowder
162	Iron Range	230	Anthony Lagoon	269	Norman River
189	Bonaparte Archipel.	231	Borroloola	270	Croydon
190	Forrest River	232	Mornington Island	271	Herbert River
191	Wyndham	233	Normanton	272	Townsville
192	Daly River	234	Mitchell River	292	North West Cape

293	Dampier	382	Mundubbera	484	Leigh Creek
294	Port Hedland	383	Maryborough	485	Broken Hill
295	Oakover River	399	Edel Land	486	Wilcannia
296	Lake Dora	400	Murchison River	487	Bourke
297	Percival Lakes	401	Nicholson Range	488	Nyngan
298	Balgo	402	Meekatharra	489	Narrabri
299	Lake Dennis	403	Wiluna	490	Armidale
300	The Granites	404	Lake Wells	491	Port Macquarie
301	Barrow Creek	405	Great Victoria	494	Lord Howe Island
302	Hatches Creek	406	Warburton	508	Bunbury
303	Austral Downs	407	Tomkinson Ranges	509	Narrogin
304	Mount Isa	408	Musgrave Ranges	510	Lake Grace
305	Cloncurry	409	Ernabella	511	Esperance
306	Richmond	410	Oodnadatta	512	Balladonia
307	Hughenden	411	Macumba Creek	513	Point Culver
308	Charters Towers	412	Warburton Creek	514	Eyre
309	Mackay	413	Cooper Creek	517	Ceduna
310	Percy Isles	414	Eromanga	518	Eyre Peninsula
327	Ningaloo	415	Quilpie	519	Whyalla
328	Exmouth Gulf	416	Charleville	520	Peterborough
329	Tome Price	417	Roma	521	Lake Victoria
330	Newman	418	Dalby	522	Menindee
331	Jigalong	419	Brisbane	523	Ivanhoe
332	Lake Disappointment	435	Houtman Abrolhos	524	Lake Cargelligo
333	Traeger Hills	436	Geraldton	525	Bathurst
334	Gibson Desert	437	Mount Magnet	526	Sydney
335	Lake Mackay	438	Lake Barlee	527	Myall Lakes
336	Haast Bluff	439	Leonora	544	Cape Leeuwin
337	Alice Springs	440	Laverton	545	Albany
338	Plenty River	441	Rason Lake	546	Cheyne Bay
339	Glenormiston	442	Jubilee Lake	554	Port Lincoln
340	Boullia	443	Serpentine Lakes	555	Kangaroo Island
341	Diamantina Lakes	444	Lake Maurice	556	Adelaide
342	Winton	445	Lake Maramangye	557	Renmark
343	Longreach	446	Cooper Pedy	558	Mildura
344	Clermont	447	Lake Eyre	559	Deniliquin
345	Emerald	448	Lake Blanche	560	Griffith
346	Rockhampton	449	Cameron Corner	561	Canberra
363	Carnarvon	450	Tibooburra	562	Wollongong
364	Gascoyne Junction	451	Cunnamulla	591	South Kangaroo I.
365	Yinnietharra	452	Brewarrina	592	Coorong
366	Peak Hill	453	Moree	593	Mount Gambier
367	Canning Gap	454	Inverell	594	Ballarat
368	Glenayle	455	Lismore	595	Melbourne
369	Gibson Desert	462	Norfolk Island	596	Albury
370	Rawlinson Range	472	Perth	597	Cooma
371	Docker River	473	Northam	598	Narooma
372	Ayers Rock	474	Merredin	629	Portland
373	Erdunda	475	Kalgoorlie	630	Warrnambool
374	Finke	476	Cundeelee	631	Wilsons Promontory
375	Simpson Desert	477	Rawlinna	632	Bass Strait
376	Birdsville	478	Nullarbor	667	NW Tasmania
377	Diamantina River	479	Forrest	668	Launceston
378	Windorah	480	Cook	669	NE Tasmania
379	Blackall	481	Immarna	703	SW Tasmania
380	Tambo	482	Tarcoola	704	Hobart
381	Taroom	483	Woomera	705	SE Tasmania

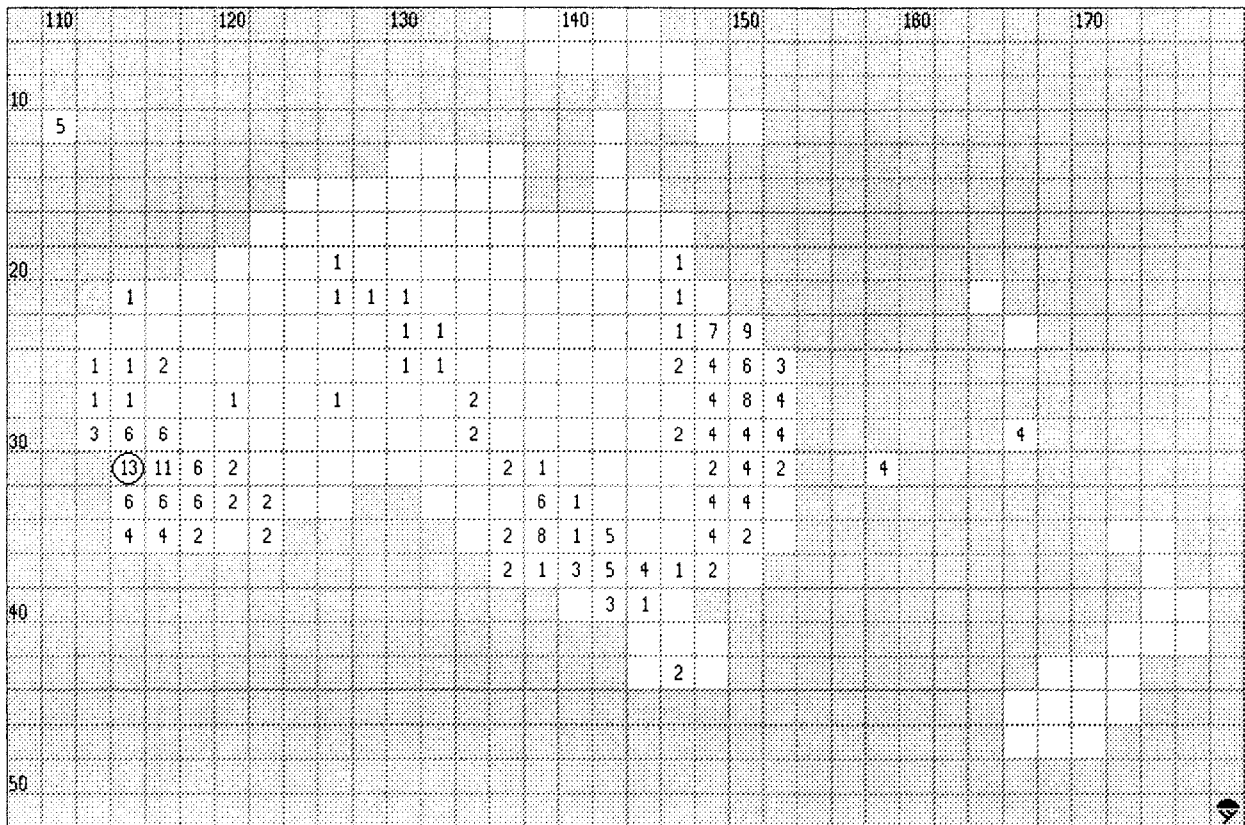
APPENDIX 8.2 MAP OF ENDEMISM VALUES BASED ON THE 47 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.

Highest endemism values are assigned to grid squares 472 (Perth, south-western WA), 110 (Christmas Island) and 520 (Peterborough, SA).



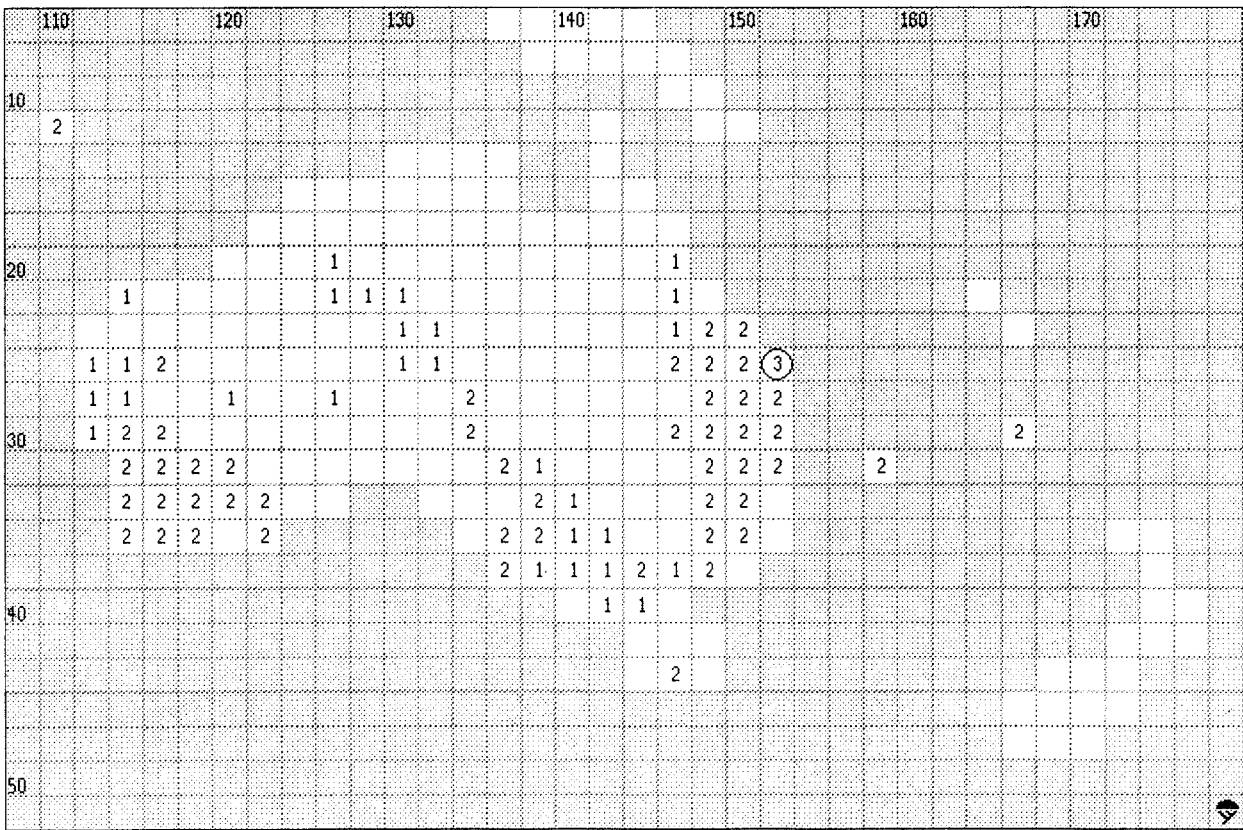
APPENDIX 8.3 MAP OF ROOT WEIGHT VALUES BASED ON THE 47 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.

Sequence of highest root weight values is found in grid squares 472 (Perth, south-western WA), 473 (Northam, southern WA), and 346 (Rockhampton, QLD).



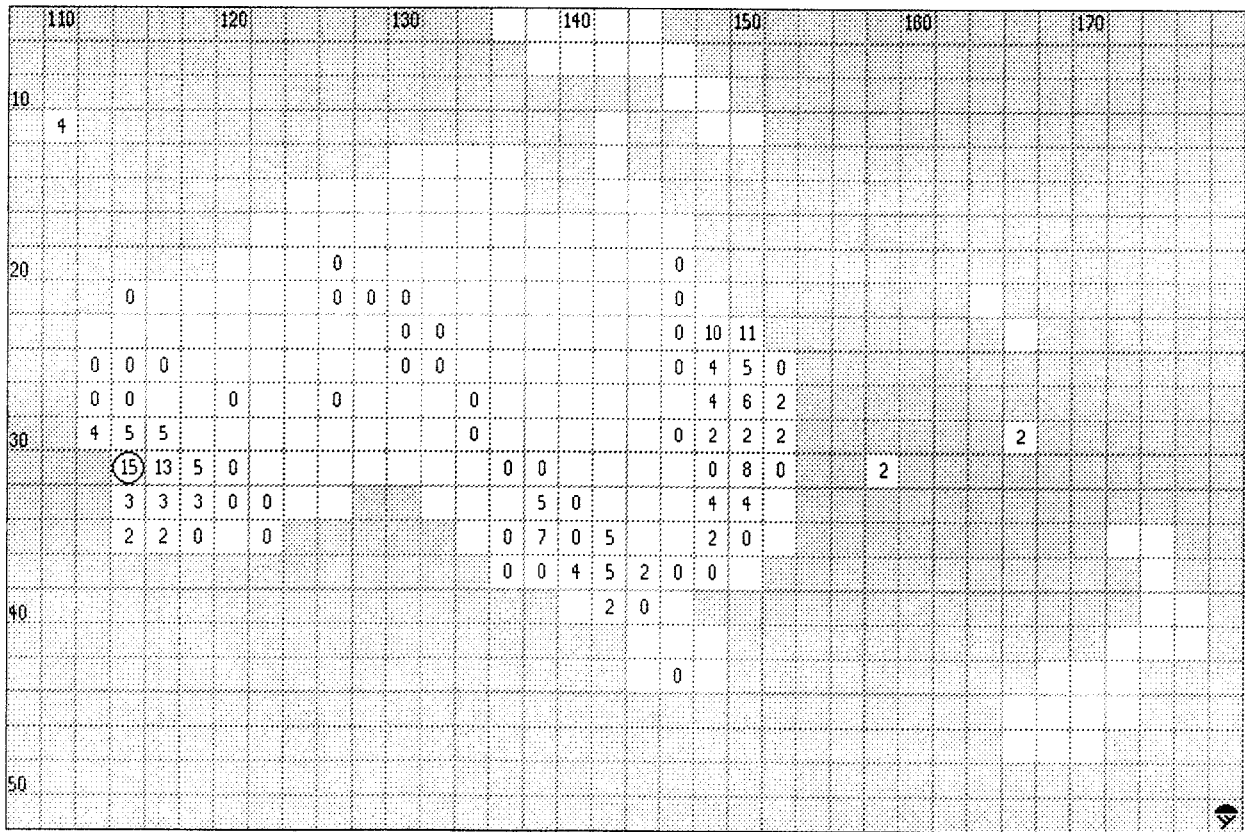
APPENDIX 8.4 MAP OF MEAN ROOT WEIGHT VALUES BASED ON THE 47 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.

Highest mean root weight value is assigned to grid square 383 (Maryborough, QLD).



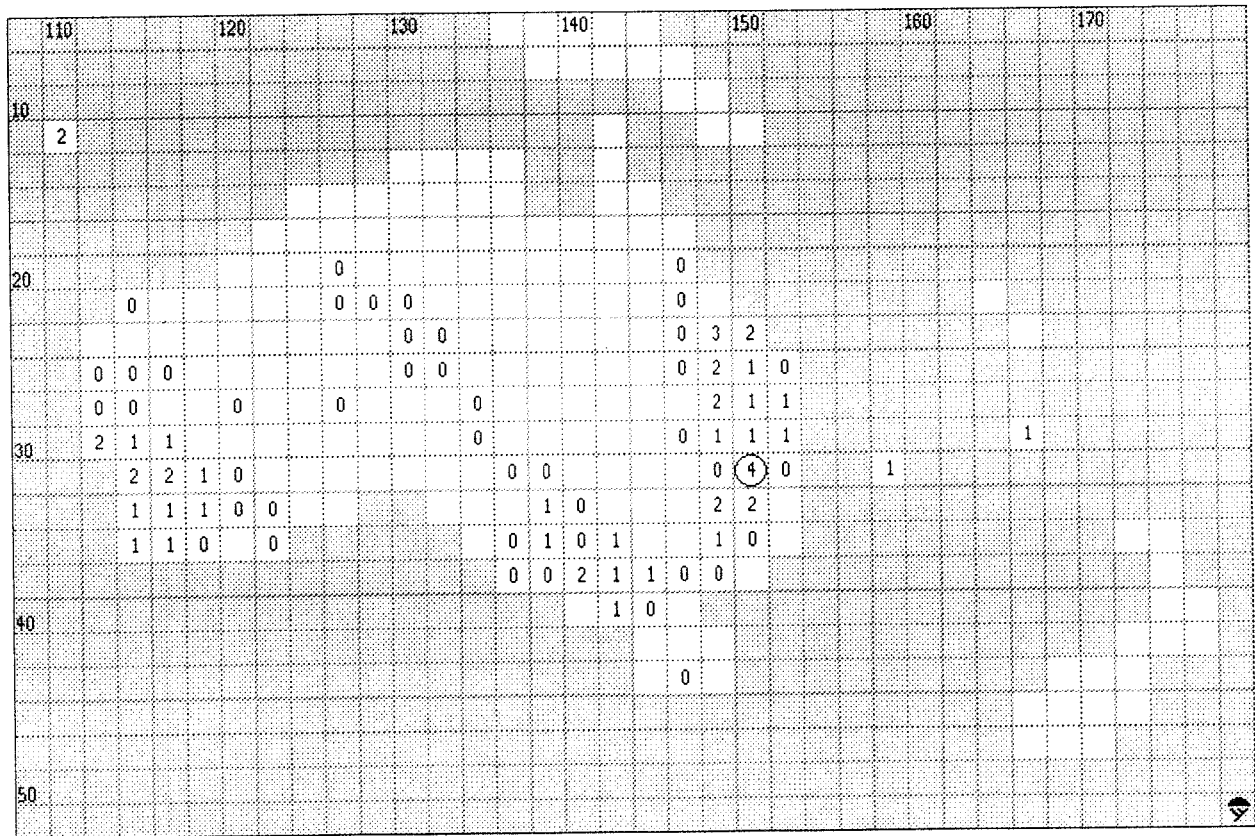
APPENDIX 8.5 MAP OF HIGHER TAXON WEIGHTINGS BASED ON THE 47 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.

Sequence of highest higher taxon weight values is found in grid squares 472 (Perth, southern WA), 473 (Northam, southern WA), 346 (Rockhampton, QLD) and 345 (Emerald, QLD). Note that grid squares represented by a single species or geographically discrete population will automatically be assigned zero values by this algorithm.



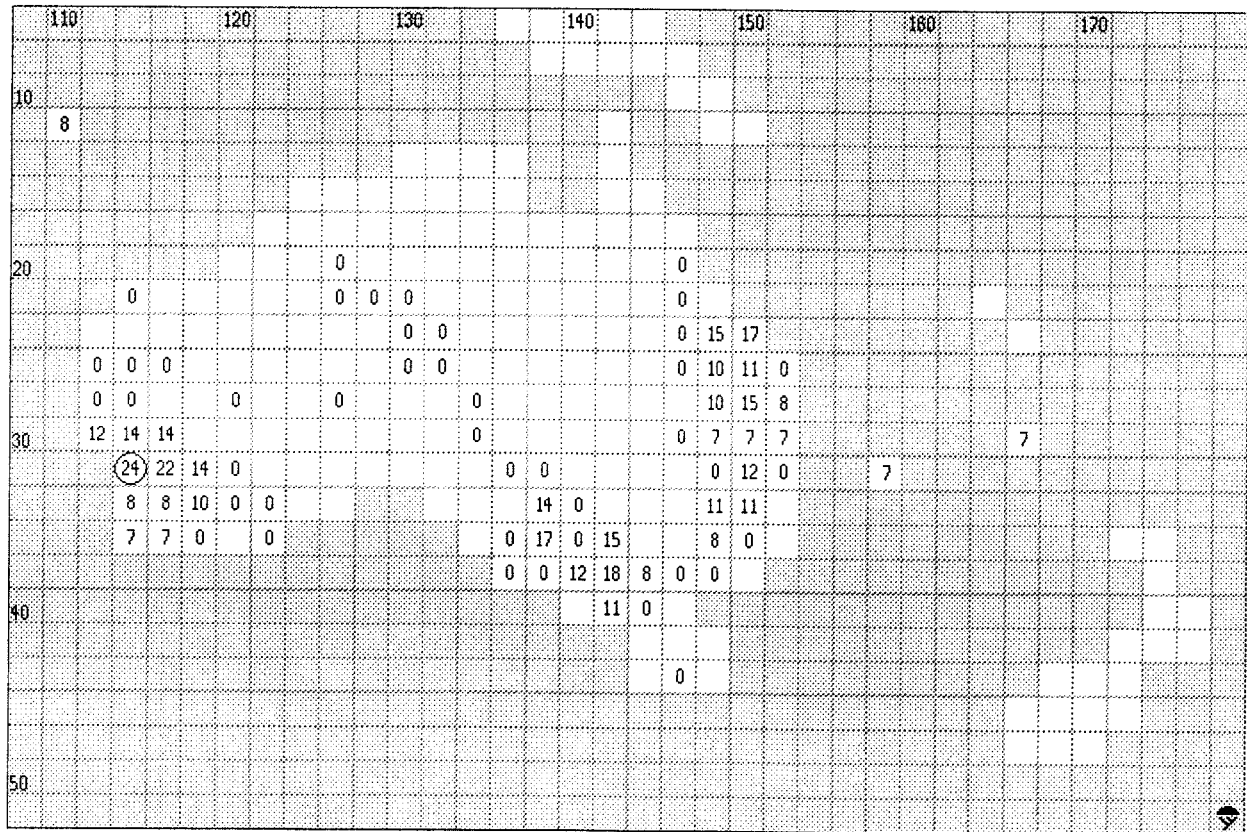
**APPENDIX 8.6 MAP OF MEAN HIGHER TAXON WEIGHTINGS BASED ON THE 47
 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN
 REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.**

Highest mean higher taxon values are found in grid squares 490 (Armidale, NSW), 345 (Emerald, QLD) and 346 (Rockhampton, QLD).



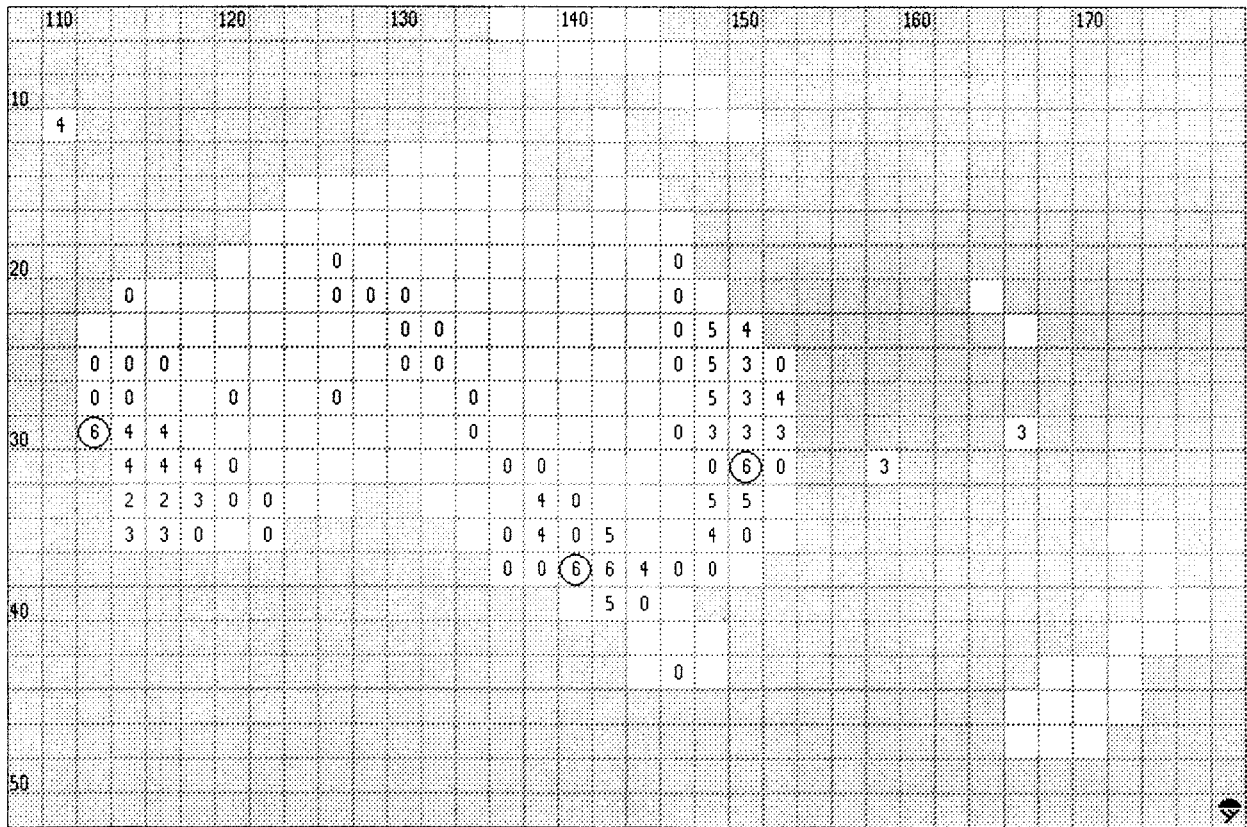
APPENDIX 8.7 MAP OF SPAN (UNROOTED-SUBTREE) WEIGHT VALUES BASED ON THE 47 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.

Highest span weight values are found in grid squares 472 (Perth, southern WA), 473 (Northam, WA), 594 (Ballarat, VIC) and 556 (Adelaide, SA).



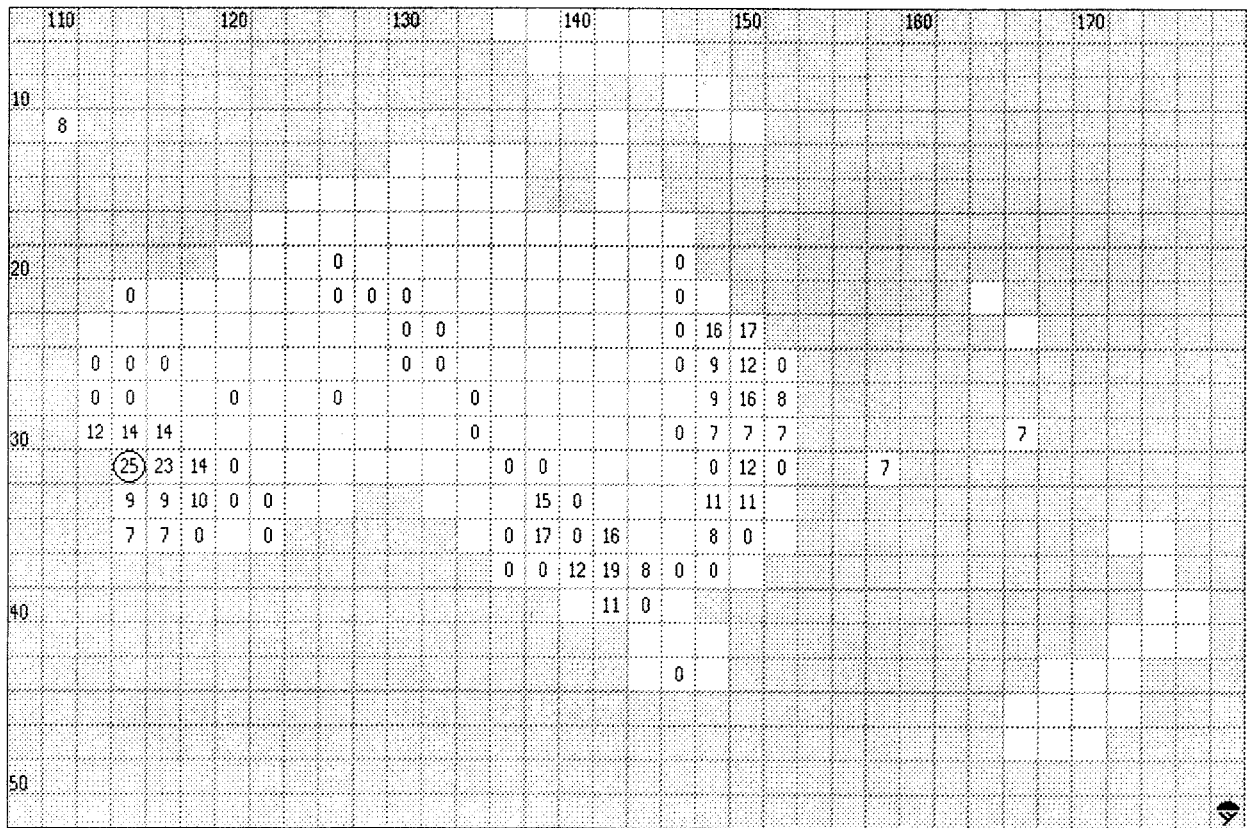
APPENDIX 8.8 MAP OF MEAN SPAN (UNROOTED-SUBTREE) WEIGHT VALUES BASED ON THE 47 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.

Highest mean span weight values are found in grid squares 435 (Houtman Abrolhos, WA), 593 (Mount Gambier, SA), 490 (Armidale, NSW) and 594 (Ballarat, VIC).



APPENDIX 8.9. MAP OF DISPERSION WEIGHT VALUES BASED ON THE 47 TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.

Sequence of highest dispersion weight values is found in grid squares 472 (Perth, southern WA) and 473 (Northam, WA).



**APPENDIX 8.10 MAP OF MEAN DISPERSION WEIGHT VALUES BASED ON THE 47
TERRESTRIAL SPECIES OR GEOGRAPHICALLY DISCRETE POPULATIONS OF AUSTRALIAN
REPTILES PROPOSED AS ENDANGERED OR VULNERABLE.**

The highest mean dispersion weight value is found in grid square 594 (Ballarat, VIC), 435 (Houtman Abrolhos, WA), 593 (Mt Gambier, SA) and 490 (Armidale, NSW).

